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Phosphatase activity as a parameter for assessment of the rhizodegradation potential of poplar clones: Greenhouse doseresponse experiment of phytoremediation of oil contaminated soil

Andrej Pilipović¹, Saša Orlović¹, Marina Katanić¹, Jelica Simeunović², Saša Pekeč¹, Milan Matavuly^{3*}

¹University of Novi Sad, Institute of Lowland Forestry and Environment, Antona Čehova 13, 21000 Novi Sad, Republic of Serbia ²University of Novi Sad, Faculty of Science, Department of Biology and Ecology, Trg Dositeja Obradovica 2, Novi Sad, Republic of Serbia

³Faculty of Pharmacy – European University, Novi Sad, Trg Mladenaca 5, 21000 Novi Sad, Republic of Serbia

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Summary. During rhizo-degradation, certain microorganisms are capable of breaking down hazardous pollutants such as naphtha and its derivatives or other xenobiotics into non-toxic products. Some plant clones with improved plant-microbe interactions have been selected which should display enhanced ability to secrete exudates which stimulate microbial activity in the substrate where phyto-degradation occurs. Metabolic processes performed by microorganisms of the rhizosphere, and in particular enzymatically catalyzed bio-degradation, form the basis of bioconversion processes for persistent and toxic pollutants into non-toxic or less toxic byproducts. Such bio-conversion is in fact the essence of both spontaneous and controlled bio-remediation processes. Thus, these processes can be monitored by measuring the activity levels of highly specific or less specific enzymes. One universal group of enzymes present in microorganisms involved in bioconversion are phosphomonoester-hydrolases: including alkaline, acid and neutral phosphatase enzymes. Phosphomonoester-hydrolase activity is a good indicator of organic load in natural freshwater, as well as during bio-remediation processes that occur in artificial and natural ecosystems. The aim of this study was to investigate the rhizo-degradation potential of different poplar (Popullus spp.) clones in oil contaminated soil, and to measure phosphomonoester-hydrolase activity in rhizosphere microflora as an indicator of rhizo-remediation processes. The effect of crude naphtha on the number of aerobic heterotrophic, facultative oligotrophic and naphtha-oxidizing bacteria, along with the phosphomonoester-hydrolase activity present in rhizosphere microflora, was studied using a greenhouse dose-response experiment. Cuttings from three poplar clones were planted into pots with soil loaded with 6 dosage levels of oil contaminated dry soil ranging from 0-6174 mg/kg of crude naphtha. Six months after planting, both the number of bacteria and phosphatase activity in the rhizosphere zone were increased in low contaminated soil (5%), but significantly decreased with increasing of oil contamination (50%) in a dose-response manner, to a maximum 10-fold reduction in 100% naphtha contaminated soil (vs. 0% contaminated soil controls). No differences were observed between different poplar clones with respect to support of rhizosphere bacterial growth. The relationship between phosphatase activity and the abundance of different groups of bacteria suggest that this phosphatase activity is associated with a specific group of bacteria.

Keywords: indicator, oil polluted soil, phosphatase activity, poplar clones, rhizodegradation.

INTRODUCTION

Soil contaminated with oil and its derivatives represents a significant environmental problem, particularly due to the ever increasing use of fossil fuel based transportation, machinery and lubricants, as well as accidental oil spills and other environmental disasters (Petrovicy et al. 1985, 2004; Vidyikant et al. 2007). Everywhere they are processed, transported, moved, stored or used, petrochemical products contaminate the surrounding soil. In particular, the oil and natural gas industry utilize a series of complex technological processes that are potential hot-spots of environmental contamination (Rončević 2002; Yovanovicy et al. 2003, 2005, 2006; Popov et al. 2008; Marjanovic et al. 2009). Recent studies have demonstrated the persistence of these compounds in ecosystems, and the possibility of their migration in the soil and subsequent release into the groundwater; spreading their impact to a much greater area beyond the spill center (Petrovic et al. 2003; Petrović et al. 2001, 2004).

During the 1999 NATO bombing, the oil refinery of Novi Sad suffered significant damage where, of 73,569 tons of oil and oil products, approximately 90% burned, 9.9% was spilled on the soil surface, and around 0.1% was leaked directly into the Danube river. The total surface area of gasoline contaminated soil was approximately 8,500 m², while 51,000 m² was polluted with crude naphtha, and 35,000 m² was contaminated with oil derivatives such as diesel and kerosene (Nježić and Ačanski 2009). Thus, oil leakage is a significant environmental problem due to the proximity of the Danube and the Renney wells, which supply the city of Novi Sad with water (Petrovic et al. 2003, Petrović et al. 2004). Because of this, soil cleanup technology is essential to prevent the further spread of pollution and penetration into groundwater and surface waterways.

The processes of soil remediation using mechanical and physical-chemical methods is very expensive, ranging between \$100-\$1,500 per ton of soil, depending on the purification/processing techniques utilized (Schnoor 1995).

In contrast to the above mentioned methods, an alternative may be the application of phytoremediation techniques which can cost ten-fold less. Phytoremediation is the use of plants and their associated microorganisms for environmental cleanup (Salt et al. 1995; Raskin et al. 1997). Phytoremediation technology uses the natural processes by which plants and rhizosphere microorganisms decompose and deposit organic and inorganic pollutants (Pilon-Smits 2005).

The decomposition of organic compounds via the activity of microorganisms of the plant rhizospere is called rhizodegradation. Rhizodegradation, also known as phytostimulation, is the degradation of contaminants in the rhizosphere (the area of soil surrounding plant roots) by means of microbial activity, coupled with secreted plant root exudates containing carbon and nitrogen compounds which provide the microorganisms with additional nutrients. Microorganisms such as yeast, fungi or bacteria consume contaminants as a source of energy and nutrition (Matavulj et al. 1983; Katanić et al. 2013). During this process of biodegradation, certain microorganisms are capable of breaking down hazardous pollutants, such as naphtha and its derivatives and other xenobiotics into nontoxic and harmless products. It is possible to select clones with improved plant-microbe interactions. Such plants would have enhanced ability to secrete natural substances which stimulate microbial activity (Maqbool et al. 2013; Ouvrard et al. 2014).

The term rhizosphere was originally described in leguminous plants by Hiltner 1904 and according to Curl and Truelove (1986) represents a zone of increased numbers of microorganisms and microbial activity on the root surface. The rhizosphere is a perfect example of co-operation, where a consortium of different microorganisms each contribute specific enzymatic activities required for synergistic, efficient degradation of complex organic compounds that no single microorganism is capable of degrading alone (Andretson et al. 1993; Walton et al. 1994).

Rhizosphere microorganisms exudate various enzymes which enable various processes of bio-transformation in the rhizosphere, including oxidation, reduction, hydrolysis, conjugation and others (Coats 1991; Lu et al. 2009). Plants can enhance the bio-degradation of organic pollutants in their rhizosphere, in a process called phyto-stimulation (Mc-Cutcheon and Schnoor 2003). Phyto-stimulation is useful for the biodegradation of hydrophobic organic compounds, such as polychlorinated biphenyls (PCBs), polyaromatic hydrocarbons (PAHs), and other petroleum hydrocarbons, that cannot be absorbed, but can be broken down (Hutchinson et al. 2001).

The interaction of plants and microorganisms during phytostimulation is an example of mutual aid, since microorganisms create favorable conditions for the life of plants by detoxification of phytotoxic compounds, and by increasing the availability of some nutrients (Anhalt et al. 2000), while plants provide other nutrients for microorganisms through root exudates (Chappell 1997). In addition, intensive growth of plants together can increase the biodiversity of the rhizosphere and expand the range of possible biotransformations (Belden and Coats, 2004).

The metabolic processes performed by rhizophere microorganisms, such as enzymatically catalyzed biodegradation, are essential for the bioconversion of persistent and toxic pollutants into non-toxic or less toxic products: the essence of spontaneous or controlled bioremediation processes. These processes can be monitored by measuring the activity levels of highly specific or less specific enzymes. Phosphomonoester-hydrolases (including alkaline, acidic and neutral phosphatase enzymes) represent a universal group of enzymes present in rhizosphere microorganisms (Matavulj et al. 1976, 1990; Matavulj 1986, 1997; Babić et al. 2013). Experiments have shown that phosphomonoester-hydrolase enzymes are actively involved in the degradation of organic pollution in water and soil (Matavulj and Flint 1987; Song et al. 2006; Topalova et al. 2013), and are also involved in the purification of persistent, hard-to-degrade oil-based pollutants (Petrović et al. 1982, 2001; Matavulj et al. 1983; Jakšić et al. 2002). In addition, phosphomonoester-hydrolase activity has been shown to be a good indicator of organic load in natural freshwater (Matavulj et al. 1984, 1997, 2000; Matavulj 1997; Zlatković et al. 2010), as well as for bioremediation processes in artificial (Matavuly et al. 2002, 2003; Schneider and Topalova 2013) and natural ecosystems (Matavulj et al.

1989, 2007; Nemes et al. 2008; Todorova and Topalova 2013).

In light of the above, the aim of the present study was to investigate the rhizodegradation potential of different poplar (*Popullus* spp.) clones in oil-contaminated soil, and to measure phosphomonoester-hydrolase activity in rhizosphere microflora as an indicator of bioremediation.

MATERIAL AND METHODS

Experiments were conducted in terrestrial cultures of different poplar clones under semi-controlled conditions. Varying amounts of oil-contaminated soil (0%, 5%, 25%, 50%, 75%, and 100% by volume) taken from the Novi Sad oil refinery was mixed with uncontaminated soil and used as a substrate for plant growth. The concentrations of pollution per treatment are shown in Table 1. Cuttings of following poplar clones: (I) Populus nigra x maximowitizil cl.9111/93; (II) Populus deltoides cl. "Bora" and (III) Populus x euramericana cl. "Pannonia", were planted in the above prepared soil. Plants were grown from May to October, and soil samples for microbiological investigations were collected in late September (Pilipović et al. 2005). Vessels without plants were used as a control, to determine the impact of plants on the growth of microorganisms. Microbiological studies consisted of: (I) determination of the total number of heterotrophic bacteria, facultative heterotrophic bacteria and naphtha-oxidizing aerobic heterotrophs; and (II) the phosphatase (acid, neutral and alkaline phosphomonoester-hydrolase) activity of rhizosphere microorganisms. Sampling was conducted in the rhizosphere zone in four places in each pot, and samples were used for a composite sample mixture. Samples were stored in a refrigerator at +4 °C until analysis (< 1 month).

Determination of the total number of microorganisms was performed using the standard pour-plate method (Gajin et al. 1987; Matavulj et al. 1996), by growth on solid agar nutrient medium in petri plates. Determination of the number of indicator groups of saprotrophic aerobic bacteria was conducted in four replicates on different media: (I) nutrient agar - to determine the total number of aerobic heterotrophic bacteria, (II) agar naphtha - to determine the total number of oil-oxidizing bacteria (Rodina 1972) and (III) ten-fold diluted agar nutrient medium - for determination of facultative-oligotrophic bacteria, the dominant group in habitats poor in easy-to-degrade organic substrates (Ishida et al. 1982; Gayin et al. 2003; Campbell et al. 2013). Previous studies suggest that oligotrophic bacteria are predominant in soil, even on carbon-rich sites such as root surfaces (Ohta and Hattori 1983).

Petri plates were incubated after inoculation at 22 °C for 7 days. After incubation, colonies were counted and total colony number expressed per 1 g of sample.

Phosphatase activity was determined using p-nitrophenylphosphate as a substrate (Matavulj and Flint 1987). Samples were transferred to spectrophotometric plates in three wells containing TTA buffer at pH 5, 7 or 9 respectively, and p-NPP (para-nitrophenylphosphate) was added. Spectrophotometer plates were incubated in a water bath at 30 °C for 60 minutes. After incubation, the reaction was stopped by addition of 10M NaOH and the yellow-colored p-nitrophenol was developed. Samples were read in a spectrophotometer (Thermo Multiskan EX Labsystens), and on the basis of these results phosphatase activity was calculated. Results were expressed using the phosphatase activity index (PAI) – the mean activity values of acid, neutral and alkaline phosphatase (Matavulj 1986, 1990).

Total petroleum hydrocarbon (TPH) and mineral oil content in soil samples were determined according to Škunca–Milovanović et al. (1990) using quantitative IR spectrophotometry. Soil porosity and air capacity was measured according to Bošnjak et al. (1997), while C/N ratio was determined by analysis of carbon and nitrogen using a CHN Elementar III analyzer, which works on the principle of oxidation and reduction of elements at temperatures of 900 °C and 500 °C, respectively. Calculations and statistical analysis of the data was performed using Microsoft Office Excel. The Excel Analysis tool pack was used for regression analysis.

RESULTS

In order to measure the effect of oil contaimination on rhizosphere bacterial growth, the total number of heterotrophic, facultative heterotrophic and naphtha-oxidizing aerobic heterotrophic bacteria were determined. As can be seen in Table 2, increased soil contamination by naphtha is associated with a reduction in the number of facultative oligotrophic bacteria in the rhizospheres of the tested poplar

Table 1. Concentrations of oil pollutants, total petroleum hydrocarbons (TPH), mineral oil and other characteristics of tested soils (g/kg soil).

| % of contami- nated soil | Moisture (%) | TPH (mg/kg) | Mineral oil (mg/kg) | C/N ratio | Porosity (%) | Air capacity (%vol) |
|-----------------------------|--------------|----------------|------------------------|--------------|-----------------|------------------------|
| 0% | 3.25 | 20 | 0 | 79.98 | 51.91 | 44.69 |
| 5% | 0.26 | 578 | 254 | 72.16 | 46.01 | 30.88 |
| 25% | 0.28 | 2812 | 1571 | 36.96 | 42.06 | 29.24 |
| 50% | 0.44 | 6046 | 3769 | 23.69 | 40.63 | 28.32 |
| 75% | 0.43 | 9311 | 5370 | 17.71 | 44.47 | 32.69 |
| 100% | 0.42 | 10841 | 6174 | 14.31 | 46.08 | 36.34 |

| | | | Phosphatase activit | | |
|------------------------------|----------|------------------------------------|-----------------------------------|------------------------------|--------------------------|
| % of contaminated soil added | Clone | Facultative oligotrophs (cfu/g) | Aerobe heterotrophs (cfu/g) | Naphtha-oxidizing (cfu/g) | PAI (µmol pNP/s/ dm³) |
| %0 | Pannonia | 1.29E+06 | 2.00E+06 | 1.21E+06 | 9.80 |
| | 9111/93 | 1.17E+06 | 2.75E+06 | 1.62E+06 | 6.95 |
| | Bora | 1.26E+06 | 2.69E+06 | 1.27E+06 | 6.16 |
| | x̄clones | 1.24E+06 | 2.48E+06 | 1.37E+06 | 7.64 |
| | Control | 1.09E+06 | 1.71E+06 | 1.06E+06 | 7.70 |
| 5% | Pannonia | 1.95E+05 | 1.47E+06 | 1.03E+06 | 14.72 |
| | 9111/93 | 1.94E+06 | 2.29E+06 | 1.34E+06 | 12.55 |
| | Bora | 1.38E+06 | 2.67E+06 | 1.85E+06 | 6.05 |
| | x̄clones | 1.17E+06 | 2.14E+06 | 1.41E+06 | 11.11 |
| | Control | 1.42E+06 | 1.67E+06 | 1.65E+06 | 7.93 |
| | Pannonia | 1.86E+06 | 1.97E+06 | 3.95E+05 | 5.05 |
| | 9111/93 | 9.78E+05 | 2.29E+06 | 7.23E+05 | 3.48 |
| 25% | Bora | 7.25E+05 | 2.11E+06 | 7.50E+05 | 8.26 |
| N | x̄clones | 1.19E+06 | 2.12E+06 | 6.23E+05 | 5.60 |
| | Control | 1.51E+06 | 1.07E+06 | 9.43E+05 | 8.11 |
| | Pannonia | 8.55E+05 | 1.09E+06 | 8.70E+05 | 0.68 |
| | 9111/93 | 5.92E+05 | 6.20E+05 | 2.56E+05 | 6.10 |
| 50% | Bora | 5.76E+05 | 7.22E+05 | 3.04E+05 | 5.09 |
| Ŋ | x̄clones | 6.74E+05 | 8.11E+05 | 4.77E+05 | 3.96 |
| | Control | 1.85E+06 | 7.34E+05 | 2.07E+05 | 1.21 |
| | Pannonia | 4.06E+05 | 5.96E+05 | 2.01E+05 | 4.83 |
| | 9111/93 | 4.63E+05 | 8.12E+05 | 1.76E+05 | 2.03 |
| 75% | Bora | 1.60E+05 | 6.33E+05 | 1.52E+05 | 1.11 |
| | x̄clones | 3.43E+05 | 6.80E+05 | 1,76E+05 | 2.65 |
| | Control | 1.80E+06 | 9.03E+05 | 1.76E+05 | 2.30 |
| | Pannonia | 5.32E+05 | 5.82E+05 | 2.18E+05 | 1.00 |
| .0 | 9111/93 | 1.22E+05 | 6.21E+05 | 2.30E+05 | 0.17 |
| 100% | Bora | 4.00E+05 | 3.97E+05 | 2.31E+05 | 0.91 |
| | x̄clones | 3.51E+05 | 5.33E+05 | 2.26E+05 | 0.69 |
| | Control | 1.23E+06 | 5.02E+05 | 7.61E+05 | 2.56 |

Table 2. Number of bacteria and phosphomonoester-hydrolase activity in poplar rhizospheres.

clones, as well as in control pots without plants (Table 2). With respect to total number of bacteria, an overall decreasing trend was observed. However, clones 9111/93 and "Bora", had an increased number of facultative oligotrophs at 5% oil contamination, and clone "Pannonia" had an increased number of facultative oligotrophs at 25% contamination. In control pots without plants, the number of oligotrophic bacteria was found to be higher than in vessels with plants only at 75% and 100% oil contamination.

Microbial phosphatase activity was also determined as a function of % oil contamination and expressed as the phosphatase activity index (PAI). As can be seen, there was an increase in microbial phosphatase activity in the rhizospheres of all of tested clones at 5% oil contamination, while reduced PAI values were observed at 25% oil contamination. Increased oil contamination in the growth substrate beyond 25% lead to a general decrease in microbial phosphatase activity in both plant containing pots and control pots without plants.

Results from regression analysis and coefficient of determination (see Fig. 1) revealed a positive correlation between the number of organotrophic bacteria and phosphatase activity index (PAI). The coefficient of determination (0.4475) and correlation coefficient (0.6689) were found to be significantly correlated between the analyzed parameters.

Determination of the total number of organotrophic (aerobic heterotrophs) bacteria (Fig. 1) also demonstrated that increased oil pollution in the growth substrate resulted in a reduction of organotrophic bacterial growth in the rhizospheres of the tested clones, as well as in control soil substrates without plants. At 75% oil contamination, the number of organotrophic bacteria in control pots without plants was higher than those planted with clones "Bora" and 9111/93, while the clone "Pannonia" supported the growth of less organotrophic bacteria at 5% oil contamination.

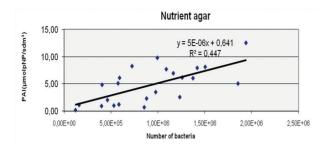


Fig. 1. Regression analysis and Coefficient of determination (R²) between the number of aerobic organotrophic bacteria and the Phosphatase activity index (PAI).

Regression analysis and the coefficient of determination (Fig. 2) revealed a positive correlation between the number of facultative oligotrophic bacteria and phosphatase activity index. The coefficient of determination (0.408) and correlation coefficient (0.6387) suggest a significant relationship between the analyzed parameters. The total number of oiloxidizing bacteria (Table 2) reduces with increasing soil pollution in both the rhizosphere of the studied poplar clones, as well as in control pots without plants.

However, at 5%, 50% and 75% oil contamination, the total number of oil-oxidizing bacteria was greater or equal to that measured in control soil samples from pots without plants. An exception to this trend is clone 9111/93, which displayed a reduced number of oil-oxidizing bacteria with increased oil pollution, while clones "Pannonia" and "Bora" supported higher numbers of naphtha-degrading bacteria at 5% ("Bora") and 50% ("Pannonia"). In general, unplanted pots supported less growth of aerobic saprotrophic bacteria vs. pots containing plants. Thus, bacterial populations were in most cases stimulated by the rhizosphere effect, in agreement with Cartmill et al. (2013).

Regression analysis and coefficient of determination between the number of oil-oxidizing bacteria (Fig. 3) and phosphatase activity revealed a positive correlation. In addition, in this case, the coefficient of determination (0.4277) and coefficient of correlation (0.6539) suggested a significant association between the analyzed parameters.

The relationship between phosphatase activity and the abundance of different groups of bacteria suggested that this phosphatase activity was associated with a specific group of bacteria. Moreover, an increased PAI/Oil-oxidizing bacteria (PAI/Naphtha) ratio was observed with increased soil pollution, which was highly evident at 50% and 75% oil contamination. These increases were particulary striking in the rhizospheres of clones 9111/93 and "Bora" treated with 50% oil contaminated soil; and in clones "Pannonia" and 9111/93 treated with 75% oil contaminated soil (Fig. 4).

DISCUSSION

The impact on the quantitative, as well as qualitative composition of indigenous microbial communities largely

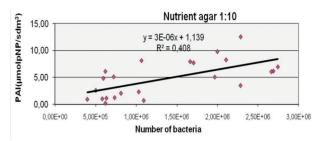


Fig. 2. Regression analysis and Coefficient of determination (R²) between the number of facultative oligotrophic bacteria and the Phosphatase activity index (PAI).

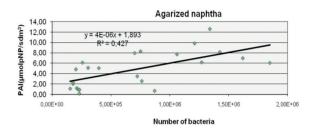


Fig. 3. Regression analysis and Coefficient of determination (R²) between the Number of naphtha-oxidizing bacteria and the Phosphatase activity index (PAI)

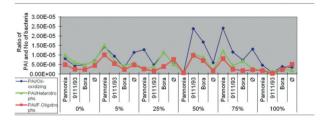


Fig. 4. The ratio of Phosphatase activity index (PAI) and number of different groups of bacteria in the rhizosphere of studied poplar clones.

depends on the particular plants in the vicinity: where different species of plants can have different effects on the local microbial community (Kirk et al. 2005). However, results of microbiological studies do not provide a complete picture of the impact of plants on microbial growth, or the degradation of oil pollution. In part this is due to the inherent complexity of the process, since interactions within the rhizosphere may greatly hamper research, understanding and subsequent conclusions (Kirk et al. 2002; Lin et al. 2002, Nie et al. 2011). Qianxin et al. (2002) found that increased oil content in soil had a greater impact on the development of Vibrio fisheri bacteria in the rhizospheres of plants than on the production of aboveground Spartinia alterniflora biomass, which can lead to underestimation of the impact of oil pollution in the soil. Wyszkowska and Wyszkowski (2008, 2010) reported that the quantitative composition and phosphatase activity of soil microorganisms depends on the dose of petroleumderived substances. Although high doses of petroleum-derived substances cause a significant increase in the number

of microorganisms in soil, the highest dose tested (10 g per kg of soil) appears to cause a decrease in the number of oligotrophic sporulating bacteria in the examined soil.

In the present study, the total number of microorganisms and their associated phosphatase activity decreased upon increasing oil pollution more than 5%, regardless of the presence or absence of plants. These results are consistent with Chekol et al. (2004), who noted a negative correlation between the number of bacteria and levels of organic contamination with Aroclor 1248 (a PCB). Similarly, Kirk et al. (2005) reported the toxic effects of diesel on the growth of mycorrhizal fungus. Also, according to Lipińska et al. (2013), urease activity was strongly inhibited by the presence of polycyclic aromatic hydrocarbons in contaminated soil; while Ohiri et al. (2013) reported strong inhibition of acid and alkaline phosphatase activity in soil contaminated with bonny light crude oil. Pérez-Leblic et al. (2012) also found that the highest values of microbial enzyme activities correspond to landfill soil areas with the lowest concentration of hydrocarbon pollutants.

In some cases the results of these investigations suggest only a weak effect due to the presence of plants on the total number of microorganisms. For example Ionescu et al. (2009) showed that the rhizospheres of tobacco, black nightshade, horseradish and willow (Salix caprea) had no significant impact on the total number of microorganisms present in PCB contaminated soil vs. contaminated soil without plants. In fact, there was no apparent correlation between the total number of heterotrophs and the number of microorganisms that degrade PCBs, which was explained by variations in exudates, soil moisture and the presence of other microorganisms related to plant species (e.g. mycorrhizal fungi). In the present study, greater or equal numbers of microorganisms were found in unplanted pots compared with pots in which poplar clones were grown. This can possibly be explained by the toxic influences of very high concentrations of pollutants on the vitality of the plants, which could result in decreased development of bacteria in the rhizosphere.

The total number of microorganisms, determined via the number of viable colonies on inoculated agar media, may not be the most reliable indicator of biodegradation potential. According to Margesin et al. (2000a), only a small fraction of the total bacteria can be isolated and grown on nutrient media. Therefore, in order to monitor the intensity of microbial metabolism it is necessary to quantify biological activity directly in the soil through the determination of soil (microbial) enzymatic activity and/or soil respiration. Determination of soil enzyme activity provides information on the metabolic activity of microorganisms present, and indicates the impact of stress on a microbial population (Matavulj, 1986, Matavulj et al. 1984; Margesin et al. 2000b). These facts could explain our observed increase in the rate of PAI and the number of oil-oxidizing bacteria. Small differences in terms of the number of bacteria and enzymatic activity

between vessels with and without plants can be explained by the fact that samples were taken at the very end of the experiment, which according to Lee et al. (2008), is a period when accessible, easy-to-degrade hydrocarbons are already consumed through microbial metabolism, resulting in a reduction in their numbers: the same investigators showed a reduction in enzymatic (dehydrogenase) activity due to the increased presence of PAHs in soil.

The decrease in microbial enzymatic activity at elevated oil pollution levels observed in this experiment can also be explained by a lack of nitrogen and phosphorus, which was confirmed earlier by Margesin et al. (Margesin et al. 2000a, Margesin 2007), who reported an increase in lipase and catalase activity after addition of different fertilizers to soil contaminated with naphtha. With respect to the utility of stimulus measures for enhanced degradation of petroleum hydrocarbons, similar results were obtained by Hawrot and Nowak (2006), who determined that additional soil fertilization coupled with periodic agitation (for aeration) had a significant impact on the degradation of pollutants in contaminated soils inoculated with microorganisms.

Soil enzyme activity can be used as an indirect indicator of the impact of additional measures on contaminant degradation. For example, Wyszkowski et al. (2006) found that increased fertilization or addition of organic matter (in the form of pine sawdust) resulted in increased microbial enzymatic activity in oil polluted soil: specifically increased dehydrogenase, urease and alkaline phosphatase activity. Thavasi et al. (2011) discovered that the addition of fertilizers and surfactants results in increased degradation of crude oil by cultured marine strains of three different types of bacteria. Onneby et al. (2006) showed that the rhizosphere of basketwork willow (Salix viminalis) grown in soil contaminated with creosote had a greater abundance of bacteria and a higher degradation of some PAHs compared with pots without plants. This phenomenon was explained by the increased presence of surface-active substances of biological origin.

Other factors also impact microbial populations and associated enzymatic/degradation activities, including watering regimens, treatment duration and the addition of stimulatory supplements. In a study conducted from 2003 to 2005, Truu et al. (2009) showed that watering willows using secondary treated wastewater increased the number of microorganisms, soil enzyme activity and nitrogen mineralization in the willow rhizosphere compared to the rhizospheres of untreated control plants. Also, an increase in the tested parameters correlated with the duration of treatment.

A possible cause of the small number of microorganisms and low degradation activity in contaminated soils could be the bioavailability of petroleum hydrocarbons in soil, which decreases with the age of the substrate. Thus, older contaminated soils are less toxic to plants and microorganisms since contaminants are more or less bonded to soil particles, making them less accessible for degradation (Tesar et al. 2002). In the present study, the contaminated soil used was the product of many years of pollution from a Novi Sad oil refinery. Therefore, a smaller number of microorganisms and reduced enzymatic activity is inevitable in treatments containing a higher proportion of polluted soil.

Taking into account results observed in the present study, in addition to plants whose exudates stimulate the growth of microorganisms, inoculation of a consortium of microorganisms capable of more rapid degradation of petroleum hydrocarbons is necessary to improve petroleum hydrocarbon degradation in contaminated soil. In addition to the degradation of contaminants, such inoculated microorganisms also improve conditions for plant development.

Interactions in the rhizosphere can be divided into two groups (Barea et al. 2005): (I) those based on dead plant material (decomposing interactions), which affect the circulation of energy and nutrients in the soil; and (II) those based on the living plant root. Both types are of great importance in agriculture and ecology. For example, Gunderson et al. (2007) inoculated the rhizosphere of a hybrid poplar with the fungus Pisolithus tinctorus, resulting in increased degradation rates and increased root biomass productivity in treated plants, as well as increased nitrogen and phosphorus absorption in plant leaves. However, inoculated microbial communities are often overgrown with autochthonous microflora already adapted to conditions in the rhizosphere. Therefore, as the best solution, isolation and laboratory amplification of indigenous microflora is imposed. Subsequent inoculation of contaminated soil and the plant rhizosphere with extra stimulation by the addition of supplements would provide an initial source of carbon for survival in the initial stages of rhizodegradation.

In general, in the present study differences in support of the growth of rhizosphere bacteria between poplar clones was not observed. These results are consistent with the findings of Jordahl et al. (1997), who reported that microorganisms in the rhizosphere of poplars were not exposed to selection pressure in terms of the proportions of specific groups of microorganisms in thte consortium. According to Tesar et al. (2002), a greater impact is due to the presence of diesel, rather than clonal specificity, while Banks et al. (2003) came to a similar conclusion using different genotypes of *Sorghum bicolor* in experiments investigating phytoremediation of soil contaminated with crude oil.

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REFERENCES

- Anderson TA, Guthrie EA, Walton BT. 1993. Bioremediation in the rhizosphere. Environmental Science and Technology. 27:2630–2636.
- Anhalt JC, Arthur EL, Anderson TA, Coats, JR. 2000. Degradation of atrazine, metolachlor, and pendimethalin in pesticide-contaminated soils: Effects of aged residues on soil respiration and plant survival. Journal of Environmental Science and Health, Part B. 35:417–438.
- Babić OB, Simeunović JB, Škrbic NZ, Kovač DJ, Svirčev ZB. 2013. Detection of phosphatase activity in aquatic and terrestrial cyanobacterial strains. Matica Srpska Journal for Natural Sciences (Zbornik Matice Srpske za prirodne nauke, Novi Sad). [accessed 5 Nov 2014];125:31–42. http://scindeks-clanci.ceon.rs/data/pdf/0352-4906/2013/0352-49061325031B.pdf.
- Banks MK, Kulakow P, Schwab AP, Chen Z, Rathbone K. 2003. Degradation of crude oil in the rhizosphere of *Sorghum bicolor*. International Journal of Phytoremediation. 5(3):225–234. doi:10.1080/16226510390255670.
- Barea JM, Pozo MJ, Azcon R, Azcon-Aguilar C. 2005. Microbial cooperation in the rhizosphere. Journal of Experimental Botany. 56(417):1761–1778.
- Belden JB, Coats, JR. 2004. Effect of grasses on herbicide fate in the soil column: Infiltration of runoff, movement, and degradation. Environmental Toxicology and Chemistry. 23:2251–2258.
- Bošnjak Đ, Dragović S, Hadžić V, Babović V, Kostić N, Burlica Č, Đorović M, Pejković M, Mihajlović TD, Stojanović S, eta al. 1977. Metode istraživanja i određivanja fizičkih svojstava zemljišta [Methods of research and determination of physical properties of soil]. JDPZ, Beograd. Serbian.
- Campbell JH, Zak JC, Jeter RM, Strauss RE. 2013. Environmental effects on distributions of culturable soil oligotrophic bacteria along an elevational gradient in the Chihuahuan Desert. Journal of Arid Environments. [accessed 5 Nov 2014];99: 41–50. http://www.sciencedirect. com/science/article/pii/S0140196313001560.
- Cartmill AD, Cartmill DL, Alarcón A. 2013. Short-term biodegradation of petroleum in planted and unplanted sandy soil. Journal of Environmental Quality. [accessed 6 Nov 2014];42, 4:1080–1085. http://www. bioportfolio.com/resources/pmarticle/764611/Short-term-biodegradation-of-petroleum-in-planted-and-unplanted-sandy-soil.html. doi:10.2134/jeq2013.03.0078.
- Chappell J. 1997. Phytoremediation of TCE using Populus, Status Report prepared for the U.S. EPA Technology Innovation Office under a National Network of Environmental Management Studies Fellowship, Compiled June - August 1997.
- Chekol T, Vough LR, Chaney RL. 2004. Phytoremediation of polychlorinated biphenyl-contaminated soils: the rhizosphere effect. Environment International. 30:799–804.
- Coats JR. 1991. Pesticide degradation mechanisms and environmental activation. In: Somasundaram L, Coats JR, editors. Pesticide Transformation Products: Fate and Significance in the Environment. American Chemical Society Symposium Series 459. Washington, DC. p. 10–29.
- Curl EA, Truelove B. 1986. The Rhizosphere. Advanced Series in Agricultural Sciences. Vol. 15. Berlin: Springer-Verlag. p. 228–240.
- Gajin S, Matavulj M, Gantar M. 1987. Osnovi mikrobiologije, nižih biljaka i gljiva. Praktikum [Fundamentals in microbiology, algae and fungi. Manual]. Novi Sad: Prirodno-matematički fakultet - Institut za biologiju. Serbian.
- Gayin S, Matavuly M, Radnovicy D, Petrovicy O, Simeunovicy Y, Borkovicy Z, Bokorov M. 2003. Microbiological and biochemical indicators of the Bachka Region water organic load of the Danube-Tisza-Danube canal network (Bechey-Bezdan stretch). Proc. VIIth Int. Symp. »Interdisciplinary Regional Research«, Hunedoara, Romania, CD-ROM, No 0318. [accessed 6 Nov 2014]; http://annals.fih.upt.ro/pdf-full/2003/ ANNALS-2003-2-04.pdf.
- Gunderson JJ, Knight JD, Van Rees KCJ. 2007. Impact of Ectomycorrhizal Colonization of Hybrid Poplar on the Remediation of Diesel Contaminated Soil. Journal of Environmental Quality. 36:927–934.

- Hawrot M, Nowak A. 2006. Effects of Different Soil Treatments on Diesel Fuel Biodegradation. Polish Journal of Environmental Studies. 15(4):643–646.
- Hutchinson SL, Schwab AP, Banks MK. 2001. Phytoremediation of aged petroleum sludge: effect of irrigation techniques and scheduling. Journal of Environmental Quality. 30(5):1516–1522.
- Ionescu M, Beranova K, Dudkova V, Kochankova L, Demnerova K, Macek T, Mackova M. 2009. Isolation and characterization of different plant associated bacteria and their potential to degrade polychlorinated biphenyls. International Biodeterioration & Biodegradation. 63(6):667–672.
- Ishida Y, Imai I, Miyagaki T, Kadota H. 1982. Growth and Uptake Kinetics of a Facultatively Oligotrophic Bacterium at Low Nutrient Concentrations. Microbial Ecology. [accessed 5 Nov 2014]; 8(1):23–32. http:// www.jstor.org/stable/4250687.
- Jakšić B, Vukić LJ, Matavulj M, Jovanović Đ. 2002. Autobioremediation potential of the Sava River ecosystem contaminated by Srpski Brod Oil refinery wastewater. Proceedings Yugoslav Conference with International Participation of the Yugoslav Association for Oil and Gas »YUNG>4P>2002«. p. 37–44.
- Jordahl JL, Foster L, Schnoor JL, Alvarez JJ. 1997. Effect of hybrid popular trees on microbial populations important to hazardous waste bioremediation. Environmental Toxicology and Chemistry.16:1318–1321.
- Katanić M, Kovačević B, Glowska N, Paoletti E, Matavulj M, Kraigher H. 2013. Naseljenost korena topola ektomikoriznim, arbuskularno mikoriznim i tamnim septiranim endofitskim gljivama [Density of ecto-mycorrhiza, arbuscular mycorrhizal and dark septated endophytic mushrooms on poplar roots]. Topola/Poplar. 191/192:17–29. Serbian.
- Kirk JL. 2005. Interactions between plants, contaminants and microorganisms during the phytoremediation of diesel contaminated soil [dissertation]. Guelph (ON): University of Guelph, Canada.
- Kirk JL, Klironomos JN, Lee H, Trevors JT. 2002. Phytotoxicity assay to assess plant species for phytoremediation of petroleum contaminated soil. Bioremediation Journal. 6:57–63.
- Kirk JL, Moutoglis P, Klironomos J, Lee H, Trevors JT. 2005. Toxicity of diesel fuel to germination, growth and colonization of *Glomus intraradices* in soil and in vitro transformed carrot root cultures. Plant and Soil. 270:23–30. doi:10.1007/s11104-004-1013-x.
- Lee S-H, Lee W-S, Lee C-H, Kim J-G. 2008. Degradation of phenanthrene and pyrene in rhizosphere of grasses and legumes. Journal of Hazardous Materials. 153:892–898.
- Lin Q, Mendelssohn IA, Suidan MT, Lee K, Venosa AD. 2002. The doseresponse relationship between No. 2 fuel oil and the growth of the salt marsh grass *Spartina alterniflora*. Marine Pollution Biology. 44(9):897–902.
- Lipińska A, Kucharski J, Wyszkowska J. 2013. Urease Activity in Soil Contaminated with Polycyclic Aromatic Hydrocarbons. Polish Journal of Environmental Studies. [accessed 17 Oct 2014];22(5):1393–1400. http://www.pjoes.com/pdf/22.5/Pol.J.Environ.Stud.Vol.22.No.5.1393-1400.pdf.
- Lu M, Zhang ZZ, Sun SS, Qiao W, Liu X. 2009. Rhizosphere enhanced remediation of petroleum contaminated soil. Huan Jing Ke Xue. [accessed 13 Oct 2014];30(12): 3703–3709. http://science.naturalnews. com/2009/341804_Rhizosphere_enhanced_remediation_of_petroleum_contaminated_soil.html.
- Maqbool F, Wang Z, Malik AH, Pervez A, Bhatti ZA. 2013. Rhizospheric biodegradation of crude oil from contaminated soil. Advanced Science Letters. [accessed 11 Oct 2014];19(9):2618–2621. http://www. ingentaconnect.com/content/asp/asl/2013/00000019/0000009/ art00020.
- Maqbool F, Wang Z, Xu Y, Zhao J, Gao D, Zhao Y-G, Bhatti ZA, Xing B. 2012. Rhizodegradation of petroleum hydrocarbons by *Sesbania cannabina* in bioaugmented soil with free and immobilized consortium. Journal of Hazardous Materials. [accessed 21 Oct 2014];237-238:262–269. http://www.sciencedirect.com/science/article/pii/ S0304389412008539.

- Margesin R, Walder G, Schinner F. 2000b. The impact of hydro-carbon remediation (diesel oil and polycyclic aromatic hydro-carbons) on enzyme activities and microbial properties of soil. Acta Biotechnologica. 20:313–333.
- Margesin R, Schinner F, editors. 2005, Manual for Soil Analysis Monitoring and Assessing Soil Bioremediation. Series Soil Biology. Vol. 5. Berlin: Springer Verlag.
- Margesin R. 2007. Alpine microorganisms: useful tools for low-temperature bioremediation. Journal of Microbiology. 45:281–285.
- Marjanovic DM, Vukcevic MM, Antonovic DG, Dimitrijevic SI, Jovanovic DM, Matavulj MN, Ristic MĐ. 2009. Heavy metals concentration in soils from parks and green areas in Belgrade, Journal of the Serbian Chemical Society. [accessed 18 Oct 2014];74(6):697–706. http://www. doiserbia.nb.rs/ft.aspx?id=0352-51390906697M.
- Matavulj M. 1986. Nespecifične fosfomonoestar-hidrolaze mikroorganizama i njihov značaj u kruženju fosfora u akvatičnim staništima [Nonspecific phosphomonoester-hydrolases of microorganisms and their importance in the cycle of phosphorus in aquatic environments] [dissertation]. Zagreb: Faculty of Science of the Zagreb University. Serbian.
- Matavulj M. 1997. Determination of water phosphatase activity new approach for biodegradable organic load of surface freshwater monitoring. Proceedings of IAAS Seminar "Future is in Aquaculture", Novi Sad, 2-9. July 1997. p. 9–11.
- Matavulj M, Palanački V, Ristić O. 1976. Fosfatazna aktivnost heterotrofnih bakterija iz rečne i morske vode. [Phosphatase acctivity of heterotrophic bacteria from the river and sea water]. Acta Biologica Iugoslavica, Ser. B, Mikrobiologija. 13(1):89–95. Serbian.
- Matavulj M, Petrović O, Bokorov M, Dalmacija B. 1982. Neke karakteristike naselja plesni u laboratorijskom uređaju za prečišćavanje otpadnih voda rafinerije nafte [Some characteristics of population of molds in laboratory pilot plant for oil refinery wastewater purification]. Vodoprivreda (Belgrade). 14(78–79):419–427. Serbian.
- Matavulj M, Petrović O, Dalmacija B, Gajin S, Stojilković S. 1983. Fosfatazna aktivnost kao pokazatelj procesa prečišćavanja otpadnih voda rafinerije nafte u laboratorijskom uređaju [Phosphatase activity as an indicator of processes of purification of oil refinery wastewater in a laboratory pilot plant]. Vodoprivreda (Belgrade). 15(82-83):165–172. Serbian.
- Matavulj M, Gajin S, Gantar M, Petrovic O, Erbeznik M, Bokorov M, Stojilkovic S. 1984. Phosphatase activity as an additional parameter of water condition estimate in some lakes of Vojvodina Province. Acta biologica lugoslavica - serija B, Mikrobiologija. 21(1):53–61.
- Matavulj M, Flint KP. 1987. A model for acid and alkaline phosphatase activity in a small pond. Microbial Ecology. 13(2):141–158.
- Matavulj M, Gajin S, Erbežnik M, Bokorov M, Petrović O. 1989. Phosphatase activity of water as a parameter of the River Tisa water monitoring. Tiscia (Szeged). [accessed 13 Oct 2014];23:29–36. http://www. bio.u-szeged.hu/ecology/tiscia/t23/t_23_4.pdf.
- Matavulj M, Bokorov M, Gajin S, Gantar M, Stojilković S, Flint KP. 1990. Phosphatase activity of water as a monitoring parameter. Water Science and Technology. 22(5):63–68.
- Matavulj M, Gajin S, Bokorov M, Radkovič D. 1996. The standardization and unification of methodology for microbial density/activity determination of samples of different humidity degree (water, mud, soil, etc.). Proceedings of International FEMS Symposium on Novel Methods and Standardisation in Microbiology, Košice, Slovak Republic, July, 1-4, 1996. p. 16 (W3).
- Matavuly M, Gayin S, Petrovicy O, Radnovicy D, Tamash I, Zeremski Y, Karaman M, Bokorov M. 1997. Determination of phosphatase activity of water – new approach for organic bio-degradable contaminants of surface water monitoring. Proceedings of 32. Conference of IAD of SIL, Wien, Österreich, Band II (Wissenschaftliche Hauptreferate). p. 137–146.
- Matavuly M, Gayin S, Petrovicy O, Radnovicy D, Simeunovicy Y. 2000. The

River Danube water quality nearby Novi Sad as affected by changes in river water flows. In: Gallé L, Körmöczi L, editors. Tiscia Monograph Series "Ecology of River Valleys". Szeged (Hungary): University of Szeged, Department of Ecology. p. 155–160.

- Matavuly M, Dalmaciya B, Yakshicy B, Vukicy L. 2002. Biodegradation of hydrocarbon pollutants from oil refinery wastewater purification plant by stimulated and enriched intrinsic microbes. IAD Limnological Reports. 34:411–419.
- Matavuly M, Dalmaciya B, Yovanovicy D. 2003. Oil refinery wastewater hydrocarbon biodegradation by enriched intrinsic micrflora. Proceedings of 41st International Petroleum Conference; 6-8 Oct. 2003; Bratislava. p. 1–7.
- Matavuly M, Dalmaciya B, Yovanovicy Dy. 2007. Biodegradation of hydrocarbons from Oil refinery wastewater by enriched intrinsic microbes. Proceedings of 3rd ICCE (International Congress of Chemistry and Environment); 18-20 Nov. 2007; Kuwait. p. 121–125
- McCutcheon SC, Schnoor JL. 2003. Overview of Phytotransformation and Control of Wastes. In: McCutcheon SC, Schnoor JL, editors. Phytoremediation: Transformation ad control of Contamints. New York: Wiley Interscience. p. 3–58.
- Nemes K, Pilipovicy A, Matavuly M, Lozanov-Crkvenovicy Z, Dalmaciya B. 2008. Dynamics of environmental phosphatase activity in the course of biomonitoring of biotransformation of oil polluted soil. Proceedings of 2nd International Scientific Conference "Remediation in environmental protection – present state and future prospects"; May 14-15, 2008; Belgrade. p. 37–44.
- Nie M, Wang Y, Yu J, Xiao M, Jiang L, Yang J, Fang C, Chen J, Li B. 2011. Understanding Plant-Microbe Interactions for Phytoremediation of Petroleum-Polluted Soil. PLoS One. 6(3):e17961. doi:10.1371/journal. pone.0017961.
- Nježić ZB, Ačanski MM. 2009. Not to be forgotten: The bombing of Novi Sad – an ecological black area. Hemijska industrija. 63(2):75–78. Serbian.
- Ohiri RC, Agha NC, Nwachukwu N. 2013. Variations in phosphatase activity of crude oil and used crankase oil polluted agricultural soil. Journal of Biology, Agriculture and Healthcare. 3(17):154–160.
- Ohta H, Hattori T. 1983. Oligotrophic bacteria on organic debris and plant roots in a paddy field soil. Soil Biology and Biochemistry. 15:1–8.
- Önneby K, Pizzul L, Granhall U. 2006. Phytoremediation of a highly creosote-contaminated soil by means of *Salix viminalis*. International Poplar Commission Working Party on Envirnomental Applications of Poplar and Willow Meeting; 18-20 May 2006; Northern Ireland. [accessed 15 Oct 2014]. http://www.fao.org/forestry/10713-04715904847dc515fe8c4d8d563ffa3c.pdf.
- Ouvrard S, Leglize P, Morel J. 2014. PAH Phytoremediation: rhizodegradation or rhizoattenuation? International Journal of Phytoremediation. 16(1):46–61.
- Pérez-Leblic MI, Turmero A, Hernández M, Hernández AJ, Pastor J, Ball AS, Rodríguez J, Arias ME. 2012. Influence of xenobiotic contaminants on landfill soil microbial activity and diversity. Journal of Environmental Management. 95(Suppl.):S285–S290.
- Petrović O, Matavulj M, Dalmacija B. 1982. Zajedničko prečišćavanje rafinerijskih i komunalnih otpadnih voda u laboratorijskom uređaju [Joint treatment of municipal and oil refinery wastewater in the labortory pilot plant]. Vodoprivreda (Belgrade). 14(78-79):341–348. Serbian.
- Petrović O, Gajin S, Matavulj M, Gantar M, Dalmacija B. 1985. Mikrobiološka ispitivanja procesa prečišćavanja otpadnih voda rafinerije nafte [Microbiological investigations of processes of oil refinery wastewater purification]. Acta Biologica Iugoslavica, Seria B, Mikrobiologija. 22(1):85–96. Serbian.
- Petrović O, Dalmacija B, Simeunović J, Radnović D, Matavulj M, Gajin S. 2001. Impact of war destructions of Novi Sad oil refinery on microbiological quality of water of wells of Ratno Ostrvo aquifer. Proceedings of 30. Conference "Water protection '01": 281-284. Serbian.
- Petrovic O, Simeunovic J, Radnovic D, Matavilj M, Gajin S, Dalmacija B. 2003. Drinking water source "Ratno Ostrvo" and oil pollution - influence of the Danube and contaminated soil on microbiological water

wells quality. Book of Abstracts of the 1st FEMS Congress of European Microbiologists; June 29 - July 3 2003; Ljubljana. P11-124, p. 409.

- Petrovicy O, Simeunovicy Y, Matavuly M, Radnovicy D, Gayin S. 2004. The naphtha contamination and impact of the River Danube infiltration belt on the microbiological water quality of wells of the Novi Sad main aquifer. IAD Limnological Reports. 35:257–263.
- Pilipović A, Nikolić N, Orlović S, Petrović N, Krstić B. 2005. Ispitivanje sposobnosti fitoremedijacije nitrata različitih genotipova roda *Populus* [Investigations of phytoremediation ability of different genotypes of genus *Populus*]. Šumarstvo/Forestry. 4:35–44. Serbian.
- Pillon-Smits E. 2005. Phytoremediation. Annual Review of Plant Biology. 56:15–39.
- Popov R, Mićović N, Jovanović Đ, Matavulj M, Antonović D. 2008. Uticaj transporta i distribucije olovnih motornih benzina kao dela "životnog ciklusa" na životnu sredinu [Effect of transportation and distribution of lead gasoline as a part of "Life cycle", on the environment]. Kvalitet. 18(1-2):103–108. Serbian.
- Qianxin L, Mendelssohn IA, Suidan MT, Lee K, Venosa AD. 2002. The doseresponse relationship between No. 2 fuel oil and the growth of the salt marsh grass, *Spartina alterniflora*. Marine Pollution Bulletin. 44(9):897–902.
- Raskin I, Smith RD, Salt DE. 1997. Phytoremediation of metals: using plants to remove pollutants from the environment. Current Opinion in Biotechnology. 8:221–226.
- Rodina AG. 1965. Metodi vodnoj mikrobiologiji [Methods in aquatic microbiology]. Nauka, Moskva. Russian.
- Rončević S. 2002. Kinetika bioremedijacionih procesa u zemljištu zagađenom naftom i derivatima nafte [The kinetics of the process of bioremediation of soil contaminated with oil and oil derivatives] [master's thesis]. Novi Sad: Prirodno-matematički fakultet, Univerzitet u Novom Sadu. Serbian.
- Salt DE, Blaylock M, Nanda Kumar PBA, Dushenkov V, Ensley BD, Chet I, Raskin I. 1995. Phytoremediation: a novel strategy for the removal of toxic metals from the environment using plants. Biotechnology. 13:468–474.
- Schneider I, Topalova Y. 2013. Enzyme activities as a tool for biological control in dairy wastewater treatment. Bulgarian Journal of Agricultural Science. 19(2):132–134.
- Schnoor JL, Licht LA, McCutcheon SC, Wolfe NL, Carreira LH. 1995. Phytoremediation of Organic and Nutrient Contaminants. Environmental Science and Technology. 29:318–323.
- Song C, Cao X, Li J, Li Q, Chen G, Zhou Y. 2006. Contributions of phosphatase and microbial activity to internal phosphorus loading and their relation to lake eutrophication. Science in China. Series D, Earth Sciences. 49(Suppl. I):102–113.
- Škunca-Milovanović S, Feliks R, Đurović B. 1990. Voda za piće, Standardne metode za ispitivanje higijenske ispravnosti. Beograd: Savezni zavod za zdravstvenu zaštitu, NIP "Privredni pregled".
- Tesar M, Reichenauer TG, Sessitsch A. 2002. Bacterial rhizosphere populations of black poplar and herbal plants to be used for phytoremediation of diesel fuel. Soil Biology & Biochemistry. 34:1883–1892.
- Thavasi R, Jayalakshmi S, Banat IM. 2011. Application of biosurfactant produced from peanut oil cake by *Lactobacillus delbrueckii* in biodegradation of crude oil. Bioresource Technology. 102:3366–3372.
- Todorova Y, Topalova Y. 2013. Short-time effect of heavy metals stress on key enzyme indicators in river sediments. Bulgarian Journal of Agricultural Science. 19(2):282–285.
- Topalova Y, Schneider I, Todorova Y. 2013. Analogous modeling of nutrient transformation in Iskar River sediments at different moisture content: microbiological and enzymological indicators. Biotechnology and Biotechnological Equipment. 27(4):3923–3931.
- Truu M, Truu J, Heinsoo K. 2009. Changes in soil microbial community under willow coppice: The effect of irrigation with secondary-treated municipal wastewater. Ecological Engineering. 35(6):1011–1020.
- Walton BT, Hoylman AM, Perez MM, Anderson TA, Johnson TR, Guthrie EA, Christman RF. 1994. Rhizosphere microbial communities as a plant defense against toxic substances in soils. In: Anderson TA, Coats JR,

editors. Bioremediation Through Rhizosphere Technology. Washington, DC: American Chemical Society. p. 82–92.

- Wyszkowska J, Kucharski M, Kucharski J. 2006. Application of the Activity of Soil Enzymes in the Evaluation of Soil Contamination by Diesel Oil. Polish Journal of Environmenal Studies. 15(3):501–506.
- Wyszkowska J, Wyszkowski M. 2008. Influence of petroleum-derived substances on number of oligotrophic and copiotrophic bacteria in soil. American-Eurasian Journal of Sustainable Agriculture. [accessed 10 Oct 2014];2(2):172–179. http://www.aensiweb.com/old/ aejsa/2008/172-179.pdf.
- Wyszkowska J, Wyszkowski M. 2010. Activity of soil dehydrogenases, urease, and acid and alkaline phosphatases in soil polluted with petroleum. Journal of Toxicology and Environ Health, Part A. 73(17-18):1202–1210. doi:10.1080/15287394.2010.492004.
- Vidyikant P, Yovanovicy Dy, Matavuly M, Dragin S. 2007. Ecological consequences of lead petrol use along the motorway Belgrade – Novi Sad (Serbia). Proceedings of 3rd ICCE (International Congress of Chemistry and Environment); 18-20 Nov. 2007; Kuwait. p. 486–489.

- Yovanovicy D, Vidyikant P, Vuykovicy I, Matavuly M. 2003. Lead petrol use and environmental impact along the motorway Novi Sad – Belgrade. Proceedings of 41st International Petroleum Conference; 6-8 Oct. 2003; Bratislava. p. 7–12.
- Yovanovicy Dy, Kevreshan Zh, Matavuly M. 2005. Lead petrol use consequences on the cabbage production along the motorway in Futog. Annals of the Faculty of Engineering Hunedoara (International Journal of Ingeneering). 3(1):63–68.
- Yovanovicy Dy, Kevreshan Zh, Matavuly M. 2006. The consequences of lead petrol use on lead content in cabbage along the motorway. Petroleum and Coal. [accessed 15 Oct 2014]; 48(2):43–46. https:// www.researchgate.net/publication/26500021_the_consequences_ of_lead_petrol_use_on_lead_content_in_cabbage_along_the_motorway
- Zlatković Ś, Šabic D, Milinčić M, Knežević-Vukčević J, Stanković S. 2010. Geographical and biological analysis of the water quality of Bovan Lake, Serbia. Archives of Biological Sciences. 62(4):1083–1089.